# **Identification of Metalliferous Ecotypes of** *Cistus ladanifer* **L. using RAPD Markers**

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- Z. Naturforsch. 60 c. 229-235 (2005)

The genetic diversity of *Cistus ladanifer* ssp. *ladanifer* (Cistaceae) growing on ultramafic and non-ultramafic (basic and schists) soils in the NE of Portugal was studied in order to identify molecular markers that could distinguish the metal-tolerant ecotypes of this species. Random Amplified Polymorphic DNA (RAPD) markers were used in order to estimate genetic variation and differences between populations. The RAPD dataset was analysed by means of a cluster analysis and an analysis of molecular variance (AMOVA). Our results indicate a significant partitioning of molecular variance between ultramafic and non-ultramafic populations of *Cistus ladanifer*, although the highest percentage of this variance was found at the intra-population level. Mantel's test showed no relationship between inter-population genetic and geographic distances. A series of RAPD bands that could be related to heavy metal tolerance were observed. The identification of such markers will enable the use of *Cistus ladanifer* in phytoremediation procedures.

Key words: Cistus ladanifer, RAPD Markers, Metal Tolerance

#### Introduction

Elevated concentrations of heavy metals in soils may occur as a result of natural processes (such as the weathering of ultramafic rocks) or may accumulate due to anthropogenic influences, such as atmospheric deposition from the combustion of fossil fuels, disposal of wastes such as animal manures or sewage sludge, utilization of fertilizers and agrochemicals, and metal industries or mining (Tiller, 1989; McGrath *et al.*, 2002). Existing technologies for remediation of such soils are usually based on civil engineering methods, which are *ex situ* and expensive. Phytoremediation offers a lower cost method for soil remediation and some extracted metals may be recycled for value (Chaney *et al.*, 1997).

Serpentine outcrops and mine deposits are edaphic discontinuities in mainland regions and have been defined as ecological or edaphic islands (Lefèbvre and Vernet, 1990). The evolution of tolerance to the suite of stress factors present in these islands has led to the genesis of serpentine endemics or to the differentiation of metal-tolerant ecotypes from wide-ranging species (Proctor and Woodell, 1975; Kruckeberg, 1984; Brooks, 1987; Linhart and Grant, 1996). These metal-tolerant

species or ecotypes are useful tools for performing phytoremediation procedures.

Random Amplified Polymorphic DNA (RAPD; Williams et al., 1990) analysis is a PCR-based technique that allows the amplification of DNA fragments from anonymous genomes (Hadris et al., 1992) and it has been described as a time- and cost-effective method of surveying genetic variation, at least at low taxonomic levels (Wolff and Morgan-Richards, 1999; Kjølner et al., 2004). Despite the dominant nature of RAPD markers, they have been used extensively in population genetic studies (reviewed in Otero-Arnaiz et al., 1997) and RAPD markers linked to Cd-uptake in durum wheat (Penner et al., 1995) and to Cu-tolerance in populations of Silene paradoxa (Mengoni et al., 2000) have been identified.

Cistus ladanifer L. (crimson spot rockrose) is a common perennial shrub species occurring, from sea level to an altitude of 1500 m, in the western Mediterranean region (SW of Europe and N of Morocco and Algeria). Of the three subspecies of C. ladanifer described in the Flora Ibérica (Muñoz Garmendia and Navarro, 1993), the ssp. ladanifer has a wide distribution: it prefers acid soils but has been found growing on serpentine outcrops in the

NE of Portugal, S of Spain and N of Morocco (Freitas *et al.*, 2004; Kidd *et al.*, 2004; Alados *et al.*, 1999; Ater *et al.*, 2000) and on mine deposits in SW Portugal and SW Spain (Alvarenga *et al.*, 2004; Toro-Gordillo *et al.*, 2001). This woody plant species could be interesting in phytoremediation programs because of its high growth rate and ability to produce considerable biomass in a relatively short time when conditions permit (Núñez *et al.*, 1989).

In this study, a RAPD procedure was used to identify RAPD bands exclusive to metal-tolerant ecotypes and to analyse the genetic structure and diversity of *Cistus ladanifer* ssp. *ladanifer* populations originating on soils derived from ultramafic and non-ultramafic rocks in the Trás-Os-Montes region of NE Portugal.

## **Materials and Methods**

# Population sampling

The plants used were grown from seed collected from four populations of Cistus ladanifer ssp. ladanifer in the Trás-Os-Montes region (NE Portugal); the populations were selected by Kidd et al. (2004). The four sites are of close proximity (linear distances ranged between 3 and 30 km), of similar altitude and climatic conditions, and mainly differed on the basis of the bedrock material and resulting soil properties (Table I). The soils at the four sites were derived from basic (metabasic) rocks (B), ultramafic rocks [ultrabasic (UB) and serpentine (S)] or schists (SC). The sites and their soils were described by Kidd et al. (2004). These authors showed ultramafic populations were more Cd-(S), Co-(S and UB), Mn-(S and UB) and Nitolerant (S), while the populations B and SC also showed a certain degree of tolerance to Mn (B) and Cu, Zn (SC).

### Plant material

Leaves from 24 young plants (6 per population) were used as a DNA source for RAPD analysis. These plants were part of an experiment of heavy metal tolerance carried out in hydroponic solutions by Kidd *et al.* (2004).

### DNA extraction

DNA was extracted from leaves following the method of Ziegenhagen and Scholz (1998) for DNA extraction from difficult plant species/tissues. The DNA quality was estimated by visual inspection of bands on 1% agarose gels run in TAE buffer (tris-acetate-EDTA, 1X, pH 8) and stained with ethidium bromide (0.2  $\mu$ g/ml), and DNA quantification was performed in a fluorometer Ultraspec 2000.

#### RAPD conditions

RAPD amplifications were carried out on a Hybaid Omn-E thermocycler (Hybaid Ltd., Middlesex, UK) using 25 µl reaction mixture containing 25 ng genomic DNA, 5 pmol of primer, 100 μm of each dNTP (Roche), 1× PCR buffer (10 mm Tris[tris(hydroxymethyl)aminomethane]-HCl, pH 8.3, 50 mm KCl, 1.5 mm MgCl<sub>2</sub>; Roche) and 0.5 units of Taq DNA polymerase (Roche). The reaction mixtures were overlaid with a drop of sterile mineral oil prior to PCR amplification. Since DNAtemplate concentration has been shown to be the most influential factor in RAPD reproducibility (Ellsworth et al., 1993), prior reactions with different quantities of genomic DNA were performed to determine the most appropriate quantity ensuring a clear and reproducible banding pattern. The PCR schedule was 45 cycles of 1 min at 96 °C, 1 min at 35 °C and 2 min at 72 °C, using the fastest possible transition between each temperature. A negative control was run for each primer. DNA amplification fragments were sepa-

Table I. Mean values of some soil properties of the four study sites: soils developed over basic rocks (B), serpentine rocks (S), schists (SC) or ultrabasic rocks (UB).

Population	$pH_{\rm H_2O}*$	Mg:Ca (exchangeable)	Co [μg g <sup>-1</sup> ]	Cu [μg g <sup>-1</sup> ]	$\begin{array}{c} Mn \\ [\mu g \ g^{-1}] \end{array}$	Ni [μg g <sup>-1</sup> ]	$\operatorname{Zn}$ [ $\mu g g^{-1}$ ]
B	5.6	0.1	94	253	2263	112	118
S	6.5	1.5	178	217	2503	1500	82
SC	4.9	0.6	62	57	2660	78	95
UB	6.0	0.7	94	53	1857	963	72

<sup>\*1:2.5</sup> soil:water.

rated on 1.4% agarose gels stained with ethidium bromide ( $0.2 \mu g/ml$ ) and run in TAE buffer ( $1\times$ , pH 8). To allow comparison between different gels, a lane with a DNA molecular weight marker (M XVI, 250 bp-ladder, Roche) was included in each electrophoresis. Photographs of gels were performed with a Gel Doc 2000 camera (Molecular Analyst program of Bio-Rad). Amplified bands were designated using the primer code and their size in base pairs, as proposed by Paran *et al.* (1991). Forty 10-mer primers (Operon sets A and B; Operon Technologies, Alameda, CA) were evaluated for suitability in a preliminary screening and 35 primers were selected for the final analyses.

## Data analysis

Each PCR product (band) was assumed to represent a single bi-allelic locus. Bands were scored as present (1) or absent (0), and a matrix of the different RAPD phenotypes was assembled. Bands that were monomorphic for the entire data set were not included in the analyses (Huff *et al.*, 1993).

A genetical distance matrix between every pair of individuals was created using (1–*S*) values, *S* being the Nei and Li's (1979) Similarity Index. This matrix was produced using the RAPDPLOT program (Black, 1998), and was then used to construct a Neighbour-joining (NJ; Saitou and Nei, 1987) dendrogram by means of the NEIGHBOR program (PHYLIP 3.5c software package; Felsenstein, 1993). The robustness of the RAPD dataset/dendrogram was tested by bootstrap analysis. A total of 500 NJ dendrograms was generated and the CONSENSE program (PHYLIP 3.5c software package) was used to produce a consensus tree. Trees were visualized using the TREEVIEW program (Page, 1996).

The partitioning of molecular variance between edaphic groupings of populations (ultramafic and non-ultramafic), between populations within each group and among individuals within each population was assessed by a nested analysis of molecular variance (AMOVA; Excoffier *et al.*, 1992). Distance measurements used for this analysis were based on Euclidean distances for all pairs of individuals. The significance of the estimated partition of genetic variance was tested with a nonparametric permutational procedure (10,000 random permutations). Pairwise  $\Phi_{\rm st}$  distances (Excoffier *et al.*, 1992) were calculated among populations

and their significance was tested with 3,000 permutations. Matrices of genetic distance and logarithm transformed geographic distances were compared using a Mantel's test (Mantel, 1967), significance was tested with 10,000 random permutations. All analyses were carried out using ARLEQUIN ver 2.001 (Schneider *et al.*, 2000).

#### **Results**

# RAPD banding pattern

A total of 317 bands were obtained from the RAPD procedure (mean 9.06 bands per primer). Thirty bands, which were shared by all the individuals, were not considered for the statistical analyses. Percentages of polymorphic bands ranged from 67.57 (population B) to 74.15 (population UB), while the global percentage of polymorphism was 90.54 (Table II). Fifty bands were specific to individuals, 13 exclusive to populations (3 in population B, 4 in S, 4 in SC and 2 in UB) and 18 bands shared among plants of the ultramafic (S and UB) populations. These last 31 bands were only observed in plants that showed tolerance to Cd, Co, Cu, Mn, Ni or Zn in hydroponic cultures and they could be related to this metal tolerance (Table III).

## Cluster analysis

Dissimilarity (1-S) values (S, Nei and Li's Similarity Index, 1979) ranged from 0.328 (B and SC were the most related populations) to 0.431 (highest distance appeared between the populations SC and UB). Mean values of intra-population differences showed that UB was the most diverse population [(1-S) = 0.362].

The Neighbor-joining dendrogram established two groups, the first of which comprised the ultramafic populations S and UB. The second group was composed of the non-ultramafic populations

Table II. Number of bands, and number and proportion (in parentheses) of polymorphic bands scored from the 35 primers in each population.

Population	No. of bands	No. of polymorphic bands (%)			
B	222	150 (67.57)			
S	225	160 (71.11)			
SC	228	163 (71.49)			
UB	236	175 (74.15)			
Total	317	287 (90.54)			

Table III. Obtained RAPD bands that could be related with tolerance to heavy metals in *Cistus ladanifer*.

Bands	Present in population	Possibly related with tolerance to
OPA06 <sub>2400</sub> , OPA09 <sub>1700</sub> , OPB13 <sub>1000</sub>	В	Mn
OPA02 <sub>840</sub> , OPA02 <sub>910</sub> , OPA09 <sub>2900</sub> , OPB14 <sub>2060</sub>	S	Cd, Co, Mn, Ni
$\begin{array}{c} OPA02_{1420},\ OPA03_{570},\\ OPA08_{670},\ OPA11_{1000},\\ OPA16_{650},\ OPA16_{1260},\\ OPA18_{1040},\ OPA18_{640},\\ OPB03_{1770},\ OPB03_{1370},\\ OPB03_{1190},\ OPB04_{600},\\ OPB04_{1900},\ OPB04_{1450},\\ OPB06_{1780},\ OPB10_{800},\\ OPB10_{700},\ OPB14_{970} \end{array}$	S, UB	Co, Mn
$\begin{array}{c} OPA16_{1670},OPA17_{1700},\\ OPA19_{900},OPB04_{850} \end{array}$	SC	Cu, Zn
OPA07 <sub>810</sub> , OPB19 <sub>820</sub>	UB	Mn

B and SC (Fig. 1a). Within each group, however, populations were not clearly defined, since some individuals from different populations appeared intermixed. These results were reinforced by the consensus tree, obtained from the 500 bootstrapped dendrograms (Fig. 1b).

### Partition of variance

The nested AMOVA analysis showed significant differences between the ultramafic and non-ultramafic population groups (this factor accounted for 19.13% of total molecular variance), while the percentage of variation contributed by intra-group differences was only 6.41. The highest percentage

Table V. Relationship between matrices of genetic and log-transformed geographic (*italics*) distances (expressed in km) using Mantel's correlation test (Mantel, 1967). Genetic distances are expressed as  $\Phi_{\rm st}$ . \* Significant  $\Phi_{\rm st}$  values (P < 0.05, after 3,000 permutations).

	В	S	SC	UB
В	-	0.281*	0.068*	0.251*
S	1.20	_	0.243*	0.090*
SC	1.16	1.48	_	0.247*
UB	0.58	1.11	1.25	-

of genetic diversity was found at the intra-population level (74.46%) (Table IV). Genetic distances among populations obtained from the AMOVA ( $\Phi_{\rm st}$  values between pairs of populations) were significant, showing that the smallest separation did not correspond with the populations of closest proximity (Table V). The Mantel's Test showed that there was no relationship between genetic and log-transformed geographic distances among populations (r=-0.026; P>0.05, after 10,000 random permutations).

# Discussion

To our knowledge this is the first report estimating genetic variation in populations of *C. ladanifer*. The RAPD technique was shown to be useful, since we obtained 317 different amplification products with no prior knowledge of any genome sequence of this species.

The cluster and AMOVA analyses performed on the RAPD data showed similar patterns of genetic structure in the four populations of *C. ladanifer* studied.

Table IV. Nested analysis of molecular variance (AMOVA) for 24 individuals of *Cistus ladanifer* from 4 populations, employing 287 RAPD markers. Statistics include: d.f., degrees of freedom; SSD, sum of squares; Sig., probability of obtaining a more extreme component estimate by chance alone (after 10,000 random permutations) and Percentage of Variation contributed by each component.  $\Phi_{ct}$  is the correlation of random pairs of haplotypes drawn from a group relative to the correlation of pairs of random haplotypes drawn from the whole population.  $\Phi_{sc}$  is the correlation of random pairs of haplotypes drawn from a population relative to the correlation of pairs of random haplotypes drawn from the whole group, averaged over all populations.  $\Phi_{st}$  is the correlation of random pairs of haplotypes drawn from within sub-populations relative to the correlation of pairs of random haplotypes drawn from the whole populations.

Source of Variation	d.f.	SSD	Variance components	Sig.	% of variation	Φ Statistics
Between groups	1	119.667	6.68403 Va	< 0.001	19.13	$\Phi_{ct} = 0.19129  \Phi_{sc} = 0.07928  \Phi_{st} = 0.25541$
Between populations within groups	2	78.917	2.24028 Vb	< 0.001	6.41	
Within populations	20	520.333	26.01667 Vc	< 0.001	74.46	

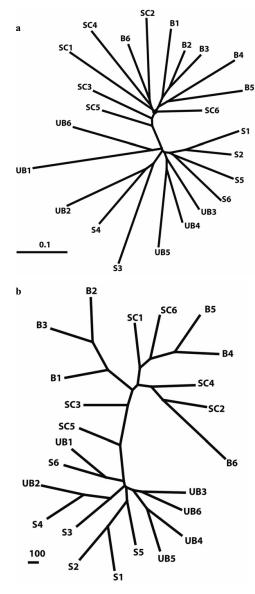


Fig. 1. (a) Neighbour-joining unrooted dendrogram (NJ; Saitou and Nei, 1987) based on Nei and Li's (1979) Similarity Index showing genetic relationships within and among populations. (b) Consensus tree based on 500 NJ dendrograms showing genetic relationships within and among populations. The populations are originated on soils derived from four rock types: basic (B), serpentine (S), schists (SC) and ultrabasic (UB). Bar indicates the genetic distance between individuals.

The Neighbor-joining dendrogram showed a clear separation between edaphic groups (ultramafic and non-ultramafic), whereas differences between populations (S/UB and B/SC) were less clear. These findings were supported by the results of the AMOVA: 19.13% of the total molecular variance was found between edaphic groups, and only 6.41% of the genetic variation was contributed by inter-population/intra-group differences.

The obtained  $\Phi_{\rm st}$  value (0.255) showed that genetic diversity is higher within *C. ladanifer* populations than between populations. This falls within the expected range for an outcrossing species (Loveless and Hamrick, 1984, and Hamrick and Godt, 1990, from protein data; Bartish *et al.*, 1999; Bussell, 1999, from RAPD data). Relatively high values of the estimates of inter-population divergence indicate a strong genetic structure. This is supported by the values of  $\Phi_{\rm st}$  distances (significantly different from 0) between pairs of *C. ladanifer* populations, indicating evidence of genetic isolation.

The lack of correlation between the matrices of genetic distances and log-transformed geographic distances in the Mantel's test indicates that geographical isolation (or isolation-by-distance) is not the main process accounting for the distribution of genetic variation among the C. ladanifer populations. Unfortunately, with the obtained RAPD data it is not possible to determine wether the origin of this population structuring could be related to phylogeopraphy of this species or if the genetic similarity between S and UB populations could be a convergence produced as result of the influence of the different stress factors present in ultramafic areas (high heavy metal content, nutrient deficiencies, low Ca/Mg ratio, high temperature, low moisture).

The existence of RAPD molecular markers distinguishing metal-tolerant ecotypes of this species would enable the use of *C. ladanifer* in phytoremediation procedures. However, further research will be necessary in order to confirm the validity of these markers.

- Alados C.-L., Navarro T., and Cabezudo B. (1999), Tolerance assessment of *Cistus ladanifer* to serpentine soils by developmental stability analysis. Plant Ecol. **143** 51–66
- Alvarenga P.-M., Araújo M.-F., and Silva J.-A. (2004), Elemental uptake and root-leaves transfer in *Cistus ladanifer* L. growing in a contaminated pyrite mining area (Aljustrel-Portugal). Water, Air Soil Poll. **152**, 81–96.
- Ater M., Lefèbvre C., Gruber W., and Meerts P. (2000), A phytogeochemical survey of the flora of ultramafic and adjacent normal soils in North Morocco. Plant and Soil 218, 127–135.
- Bartish I.-V., Jeppsson N., and Nybom H. (1999), Population genetic structure in the dioecious pioneer plant species *Hippophae rhamnoides* investigated by random amplified polymorphic DNA (RAPD) markers. Molec. Ecol. **8**, 791–802.
- Black W.-C. (1998), FORTRAN Programs for the Analysis of RAPD-PCR Markers in Populations. Colorado State University, Fort Collins, Colorado, USA.
- Brooks R.-R. (1987), Serpentine and its Vegetation: a Multidisciplinary Approach. Discorides Press, Portland, Oregon, USA.
- Bussell J.-D. (1999), The distribution of random amplified polymorphic DNA (RAPD) diversity amongst populations of *Isotoma petraea* (Lobeliaceae). Molec. Ecol. **8**, 775–789.
- Chaney R.-L., Malik M., Li Y.-M., Brown S.-L., Angle J.-S., and Baker A.-J.-M. (1997), Phytoremediation of soil metals. Cur. Opin. Biotechnol. 8, 279–284.
- Ellsworth D.-L., Rittenhouse K.-D., and Honeycutt R.-L. (1993), Artifactual variation in randomly amplified polymorphic DNA banding patterns. BioTechniques **14**, 214–217.
- Excoffier L., Smouse P.-E., and Quattro J.-M. (1992), Analysis of molecular variance inferred from metric distances among DNA haplotypes: application to human mitochondrial DNA restriction data. Genetics 131, 479–491.
- Felsenstein J. (1993), PHYLIP (Phylogeny Inference Package) version 3.5c. Distributed by the author. Department of Genetics, University of Washington, Seattle, USA.
- Freitas H., Prasad M.-N.-V., and Pratas J. (2004), Analysis of serpentinophytes from north-east Portugal for trace metal accumulation relevance to the management of mine environment. Chemosphere **54**, 1625–1642.
- Hadris H., Balick M., and Schierwater B. (1992), Applications of random amplified polymorphic DNA (RAPD) in molecular ecology. Molec. Ecol. 1, 55–63.
- Hamrick J.-L. and Godt J.-W. (1990), Allozyme diversity in plant species. In: Plant Population Genetics, Breeding and Genetic Resources (Brown A. H. D., Clegg M. T., Kahler A. L., and Weir B. S., eds.). Sinauer, Massachusetts, USA, pp. 43–63.
- Huff D.-R., Peakall R., and Smouse P.-E. (1993), RAPD variation within and among natural populations of outcrossing buffalograss [*Buchloë dactyloides* (Nutt.) Engelm.]. Theor. Appl. Genetics **86**, 927–934.
- Kidd P.-S., Díez J., and Monterroso Martínez C. (2004), Tolerance and bioaccumulation of heavy metals in five populations of *Cistus ladanifer* L. ssp. *ladanifer*. Plant Soil **258**, 189–205.

- Kjølner S., Såstad M., Taberlet P., and Brochmann C. (2004), Amplified fragment length polymorphism versus random amplified polymorphic DNA markers: clonal diversity in *Saxifraga cernua*. Molec. Ecol. **13**, 81–86.
- Kruckeberg Å.-R. (1984), California Serpentines: Flora, Vegetation, Geology, Soils and Management Problems. University of California Press, California, Berkeley, USA.
- Lefèbvre Ć. and Vernet P. (1990), Microevolutionary processes on contaminated deposits. In: Heavy Metal Tolerance in Plants: Evolutionary Aspects (Shaw J., ed.). CRC Press, Boca Raton, Florida, USA, pp. 286–297.
- Linhart Y.-B. and Grant M.-C. (1996), Evolutionary significance of local genetic differentiation in plants. Annu. Rev. Ecol. System. 27, 237–277.
- Loveless M.-D. and Hamrick J.-L. (1984), Ecological determinants of genetic structure in plant populations. Annu. Rev. Ecol. System. **15**, 65–95.
- Mantel N.-A. (1967), The detection of disease clustering and a generalized regression approach. Cancer Res. **27**, 209–220.
- McGrath S.-P., Zhao F.-J., and Lombi E. (2002), Phytoremediation of metals, metalloids, and radionuclides. Adv. Agron. **75**, 1–56.
- Mengoni A., Gonnelli C., Galardi F., Gabbrielli R., and Bazzicalupo M. (2000), Genetic diversity and heavy metal tolerance in populations of *Silene paradoxa* L. (Caryophyllaceae): a random amplified polymorphic DNA analysis. Molec. Ecol. 9, 1319–1324.
- Muñoz Garmendia F. and Navarro C. (1993), Cistaceae. In: Flora Ibérica Vol. III. (Castroviejo S., Aedo C., Cirujano S., Laínz M., Montserrat P., Morales R., Muñoz Garmendia F., Navarro C., Paiva J., and Soriano C., eds.). Real Jardín Botánico, C.S.I.C., Madrid, Spain, pp. 319–436.
- Nei M. and Li W.-H. (1979), Mathematical model for studying genetic variation in terms of restriction endonucleases. Proc. Natl. Acad. Sci. USA 76, 5269–5273.
- Núñez E., Cabeza J., and Escudero J.-C. (1989), Relación entre la biomasa de jarales y su rendimiento energetico por pirolisis. Options Méditerranéennes 3, 345–350.
- Otero-Arnaiz A., De La Cruz M., and Oyama K. (1997), El Uso de los RAPDs como marcadores moleculares en plantas. Bol. Soc. Bot. Méx. **60**, 85–117.
- Page R.-D.-M. (1996), TREEVIEW: An application to display phylogenetic trees on personal computers. Comput. Appl. Biosci. 12, 357–358.
- Paran I., Kesseli R., and Michelmore R. (1991), Identification of RFLP and RAPD markers linked to downy mildew resistance genes in lettuce, using near-isogenic lines. Genome **34**, 1021–1027.
- Penner G.-A., Clarke J., Bezte L.-J., and Leisle D. (1995), Identification of RAPD markers linked to a gene governing cadmium uptake in durum wheat. Genome **38**, 543–547.
- Proctor J. and Woodell S.-R.-J. (1975), The ecology of serpentine soils. Adv. Ecol. Res. 9, 256–347.
- Saitou N. and Nei M. (1987), The neighbour-joining method: a new method for reconstructing phylogenetic trees. Molec. Biol. Evol. 4, 406–425.
- Schneider S., Roessli, D., and Excoffier L. (2000), AR-LEQUIN ver. 2.000: A Software for Population Genetics Data Analysis. Genetics and Biometry Laboratory, University of Geneva, Switzerland.

Tiller K.-G. (1989), Heavy metals in soils and their environmental significance. Adv. Soil Sci. 9, 113–142.

Toro-Gordillo M.-C., Pinilla-Gil E., Rodríguez-González M.-A., Murciego-Murciego A., and Ostapczuk P. (2001), Potentiometric stripping analysis (PSA) for monitoring of antimony in samples of vegetation from a mining area. Fresenius J. Anal. Chem. 370, 434–437.

Williams J.-G.-K., Kubelik A.-R., Livak K.-J., Rafalski J.-A., and Tingey S.-V. (1990), DNA polymorphisms amplified by arbitrary primers are useful as genetic

markers. Nucl. Acid Res. 18, 6531-6535.

Wolff K. and Morgan-Richards M. (1999), The use of RAPD data in the analysis of population genetic structure: case studies of *Alkanna* (Boraginaceae) and *Plantago* (Plantaginaceae). In: Molecular Systematics and Plant Evolution (Hollingsworth P. M., Bateman R. M., and Gornall R. J., eds.). Taylor & Francis, London, UK, pp. 51–73.

Ziegenhagen B. and Scholz F. (1998), Methods for difficult plant species/tissues. In: Molecular Tools for Screening Biodiversity. Plants and Animals (Karp A., Isaac P. G., and Ingram D. S., eds.). Chapman & Hall,

London, UK, pp. 32–35.